

Effect of CO₂ Mitigation Incentives on the Profitability of Short-Rotation Woody Cropping of Eucalyptus amplifolia on Clay Settling Areas in Florida

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Phosphate-mined lands in central and northern Florida exceeded 69,000 ha through December 2002 and are increasing by 2,000-2,500 ha annually (Florida DEP, 2003). During phosphate mining, clays are washed from phosphate ore, and the resulting slurry is pumped into clay settling areas (CSAs). CSAs comprise about 40% of the phosphate-mined lands and can take about 15 years to dry and stabilize. While CSAs may be leased for cattle grazing for \$35-\$40 ha⁻¹ year⁻¹, they are typically left idle. Fast growing, short-rotation tree plantations provide unique opportunities to sequester carbon on CSAs and, if used as a biofuel, can reduce CO₂ emissions associated with electricity generation. *Eucalyptus amplifolia* (EA) on a CSA produced maximum sustained yields (total above-ground biomass) of 17 and 32 dry Mg ha⁻¹ year⁻¹ from initial plantings of 4,200 and 8,400 trees ha⁻¹, respectively.

Langholtz *et al.* (in press) report Land Expectation Values (LEV) for EA on CSAs in central Florida. EA production is likely to be profitable under reasonable scenarios, with LEVs ranging from \$762 to \$6,507 ha⁻¹ assuming real discount rates of 10% and 4%, respectively, establishment costs of \$1,800 ha⁻¹, planting costs of \$1,200 ha⁻¹, planting density of 8,400 tree ha⁻¹, and a stumpage price of \$20 dry Mg⁻¹. Incorporating environmental services provided by SRWC production on CSAs can improve the profitability of this system. We extend the economic optimization model of a SRWC biomass production system by Langholtz *et al.* (2006) to include an incentive for atmospheric CO₂ mitigation recognizing benefits of a) *in situ* sequestration in a mulch production scenario and b) *in situ* sequestration coupled with emissions reduction through fossil fuels displacement in a biofuels scenario. We then use these models to investigate how these incentives would influence profitability and optimal management of the SRWC production system.

METHODS

Langholtz *et al.* (2006) assessed the profitability of EA cultivation on CSAs using a modified Faustmann model as described by Medema and Lyon (1985), accounting for multiple growth stages (harvest rotations) for each coppice cycle (life of a tree) in calculating LEV. The basic Faustmann model modified by Medema and Lyon (1985) to calculate net returns given a fixed number of stages (n) is

$$LEV = \frac{\sum_{s=1}^n \left[V(t_s) \cdot e^{(-r \cdot \sum_{j=1}^s t_j)} - C_s \cdot e^{(-r \cdot \sum_{j=1}^s t_{j-1})} \right]}{1 - e^{(-r \cdot \sum_{j=1}^n t_j)}} \quad (1)$$

where $t_0=0$; n =the number of stages, s ; $V(t)$ =the growth function for stage s at time t times biomass price; r =the real discount rate; and C_s =costs of stage s at the start of the stage.

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We expand Eq. (1) to account for societal benefits of carbon (C) sequestration in the stand and subsequent decay of C in mulch products assumed to decay linearly over 5 years as

$$LEV_{mulch}(t) = \frac{\sum_{s=1}^n \left[V(t_s) \cdot e^{(-r \cdot \sum_{j=1}^s t_j)} + \left[\int_0^t \left(\frac{d}{dt} (C_s^A(t)) \right) \cdot e^{(-r \cdot t)} dt \right] \cdot e^{(-r \cdot \sum_{j=1}^s t_{j-1})} - \left[\frac{C_s^A(t)}{5} \cdot \left(\frac{1 - e^{(-r \cdot 5)}}{r} \right) \right] \cdot e^{(-r \cdot \sum_{j=1}^s t_j)} - C_s \cdot e^{(-r \cdot \sum_{j=1}^s t_{j-1})} \right]}{1 - e^{(-r \cdot \sum_{j=1}^n t_j)}} \quad (2)$$

where $C_s^A(t)$ is the value of standing aboveground C at time t for coppice stage s .

Calculation of the societal costs associated with biofuels emissions must be handled differently than Eq. (2). As described above, CO₂ emissions from sustainably produced (i.e., closed-loop) biofuels are re-sequestered in the subsequent rotation, displacing the use of fossil fuels with closed-loop biofuel resulting in no net CO₂ emissions from biomass combustion, and reducing fossil fuel emissions. However, recognizing that there are fossil fuel inputs to the cultivation, harvest, and processing of SRWCs consuming up to 10% of the energy produced by the bioenergy system (based on Heller *et al.*, 2004), 10% of the C sequestration benefit achieved at stage age t is discounted to the beginning of the stage and subtracted from the carbon benefit, resulting in

$$LEV_{biofuel}(t) = \frac{\sum_{s=1}^n \left[V(t_s) \cdot e^{(-r \cdot \sum_{j=1}^s t_j)} + \left[\int_0^t \left(\frac{d}{dt} (C_s^A(t)) \right) \cdot e^{(-r \cdot t)} dt \right] \cdot e^{(-r \cdot \sum_{j=1}^s t_{j-1})} - \left[(0.1 \cdot C_s^A(t)) \right] \cdot e^{(-r \cdot \sum_{j=1}^s t_j)} - C_s \cdot e^{(-r \cdot \sum_{j=1}^s t_{j-1})} \right]}{1 - e^{(-r \cdot \sum_{j=1}^n t_j)}} \quad (3)$$

Thus, Equations (2) and (3) are used for incorporating C externalities in mulch and biofuel production scenarios, respectively. To assess the impact of CO₂ mitigation incentives *vis-à-vis* the basic SRWC economic analysis, the model was run using the range of assumptions described by Langholtz *et al.* (2006).

RESULTS AND DISCUSSION

Under all combinations of assumptions under a fixed C price of \$5 Mg⁻¹ C, LEVs range from \$-2,789 to \$4,616 ha⁻¹ and \$-224 to \$18,121 ha⁻¹ assuming stumpage prices of \$10 and \$30 Mg⁻¹, respectively. Table 1 shows LEVs, optimum number of stages per cycle, and optimum stage lengths by C benefit scenario and stumpage price assuming a base scenario of 7% discount rate, \$1,800 ha⁻¹ site preparation cost, \$1,200 ha⁻¹ planting cost and a C price of \$5 Mg⁻¹. Increasing stumpage price decreases the optimum number of stages per cycle.

Table 1. LEV (\$/ha) and optimal number of stages and harvest age (years) for each stage by planting density (trees ha⁻¹), C benefit scenario, and biomass price.

Planting Density	NTB Scenario	\$10 dry Mg ⁻¹		\$20 dry Mg ⁻¹		\$30 dry Mg ⁻¹	
		LEV	Optimum harvest age	LEV	Optimum harvest age	LEV	Optimum harvest age
4,200	None	-2,207	3.0, 3.1, 3.2, 3.4, 4.2	-715	2.8, 2.9, 2.8, 2.7	895	2.8, 2.7, 2.6
4,200	Mulch	-2,126	3.0, 3.1, 3.2, 3.3, 3.9	-659	2.9, 2.9, 2.8, 2.7	996	2.8, 2.8, 2.6
4,200	Biofuels	-1,885	3.0, 3.0, 3.1, 3.2, 3.3	-376	2.8, 2.8, 2.9, 2.7	1,313	2.8, 2.7, 2.6
8,400	None	-798	3.3, 3.3, 3.3, 3.1	2,413	3.2, 3.1, 2.9	5,864	3.1, 3.0
8,400	Mulch	-616	3.3, 3.3, 3.3, 3.1	2,608	3.2, 3.1, 2.9	6,029	3.1, 3.0
8,400	Biofuels	-88	3.3, 3.3, 3.2, 3.0	3,197	3.2, 3.1, 2.9	6,677	3.1, 3.0

Under the base case scenario, incorporating an aboveground *in situ* C sequestration benefit of \$5 Mg⁻¹ C increases LEVs 24% and 3% to (\$946 and \$6,715 ha⁻¹) at discount rates of 10% and 4%, respectively. Recognizing the additional CO₂ mitigation benefits associated with the biofuel scenario increases, LEVs 73% and 21% (to \$1,315 and \$7,869 ha⁻¹), assuming real discount rates of 10% and 4%, respectively. In addition, the societal value of belowground C sequestration (roots + SOC at \$5 Mg⁻¹ C) is estimated at \$700 and \$1,137 ha⁻¹ at discount rates of 10% and 4%, respectively. These results emphasize both the potential for SRWCs on CSAs to mitigate atmospheric CO₂ and for CO₂ mitigation incentives to contribute to the profitability of SRWC production. The influence of stumpage price, C sequestration benefit (CO₂ mitigation scenario or C price) or discount rate (from 4% to 10%) on optimum stage lengths is less than one year.

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